

and three times that of barramundi (*Lates calcarifer*), the two other TL4 aquacultured species. The lack of association with Hg or omega-3 mirrors the lack of representation of “health” issues on webpages devoted to seafood sustainability (WebFigure 2).

Sustainability is not judged merely at a species level, but on the system that comprises the species along with vagaries of how that species is produced/harvested (Tlusty *et al.* 2012), processed, and distributed to market (Tlusty and Lagueux 2009). One difference between the Gerber *et al.* and the NGSDG data was that the latter listed a single species multiple times, representing different harvest methods and resultant sustainability scores. As an example, Pacific cod (*Gadus macrocephalus*) was present as US bottom longline (Green) and US trawled (Red). There were numerous other examples of species for which there were similar differences in sustainability status based on harvest or production method. By way of comparison, Gerber *et al.* listed each species under a single sustainability score. Such variation needs to be accounted for but will obscure the link between sustainability and human health. Care needs to be taken when selecting species for analysis, because this will substantially influence the outcome of data analyses.

The analysis presented here demonstrates that direct linkages between seafood sustainability and human health are tenuous. In this dataset, health impacts were more closely associated with TL. However, despite the inherent health “risks”, it is better to eat seafood than to not (Mozaffarian and Rimm 2006). Overall, the important assertion made by Gerber *et al.* – that eating seafood is essential both for health and for sustainability – is undeniable. Citizens in developed countries should consume more seafood because it is an efficient food source. Seaweed is the most produced aquaculture species, but little is consumed in North America. Thus, citizens in the developed world also need to select seafood choices in a fashion more aligned with global food production. Current US per capita consump-

Table 1. Simple (*r*) and partial (in bold) correlation coefficients between trophic level, sustainability score, mercury level, and omega-3 for aquaculture (above diagonal) and wild (below diagonal) species

	Trophic level	Sustainability	Mercury	Omega-3	PC1	PC2
Trophic level		−0.07 −0.08	0.51 0.44	0.72* 0.70	0.92	0.10
Sustainability	−0.16 −0.18		−0.21 −0.18	0.07 0.15	−0.14	0.91
Mercury	0.49* 0.50	−0.07 0.07		0.29 −0.11	0.71	−0.35
Omega-3	0.01 −0.15	−0.28 −0.29	0.21 0.24		0.82	−0.35
PC1	0.74	−0.48	0.78	0.47		
PC2	0.46	0.61	0.34	−0.66		

Notes: Statistically significant simple correlations are indicated by an asterisk ($P < 0.01$). The first and second principal components (PC1 and PC2) are unrotated factor patterns based on a principal components analysis (JMP 8.0, SAS Institute Inc, Cary, NC). Data are from National Geographic (2012).

tion of clams has decreased to 60% of what it was a decade ago (NFI 2012). Increasing consumption of seaweed and clams would be a basic step to take to be more sustainable in our seafood choices. A simple rule of thumb – to eat more low-trophic species – is a means to improve human and ocean health, and doing so will have positive impacts on our health and our journey to sustainability.

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- Dianto Kemmerly J. 2009. Monterey Bay Aquarium's Seafood Watch Programme. In: Ward TJ and Phillips BF (Eds). Seafood ecolabelling: principles and practice. Oxford, UK: Wiley-Blackwell.
- EPA (Environmental Protection Agency). 2012. What you need to know about mercury in fish and shellfish. http://water.epa.gov/scitech/swguidance/fish/shellfish/outreach/advice_index.cfm. Viewed 9 Aug 2012.
- Mozaffarian D and Rimm EB. 2006. Fish intake, contaminants, and human health: evaluating the risks and the benefits. *J Amer Med Assoc* **296**: 1885–99.
- NFI (National Fisheries Institute). 2012.

- Top 10 consumed seafoods. www.aboutseafood.com. Viewed 21 Sep 2012.
- National Geographic. 2012. Seafood decision guide. <http://ocean.nationalgeographic.com/ocean/take-action/impact-of-seafood/#/seafood-decision-guide/>. Viewed 8 Aug 2012.
- Tlusty MF and Lagueux K. 2009. Isolines as a new tool to assess the energy costs of the production and distribution of multiple sources of seafood. *J Clean Prod* **17**: 408–15.
- Tlusty MF, Tausig H, Taranovski T, *et al.* 2012. Refocusing seafood sustainability as a journey using the law of the minimum. *Sustainability* **4**: 2038–50.

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Does trophic level predict seafood sustainability?

We are glad that our recent paper (*Front Ecol Environ* 2012; **10**[9]: 487–493) has stimulated new approaches to communicating seafood sustainability. In his letter, Tlusty proposes that trophic level (TL) is an overlooked predictor of sustainability, and we agree that TL is an important ecological indicator in marine ecosystems. After implying that our analysis did not explicitly include TL, Tlusty used a smaller database ($n = 64$ versus $n = 362$) to suggest that (1) TL is strongly related to methylmercury levels and hence health and (2) the link

between health and sustainability metrics is tenuous. We agree with Tlusty's first hypothesis but not with his second.

First, we respectfully disagree that the link between health and sustainability is unsubstantiated. We suspect that the author's statistical power was limited given the smaller size and nature of his dataset. Notably, the National Geographic Seafood Decision Guide (NGSDG) incorporates much of the same information used in the Monterey Bay Aquarium (MBA) sustainability rankings that we used for one of our analyses, and for a largely overlapping mercury (Hg) dataset as well. Tlusty states that his database is stock-specific and that we only used one sustainability metric per species (whereas the NGSDG uses multiple). Both assertions are incorrect: we looked at multiple disparate sustainability metrics for each of our stocks and arrived at similar conclusions. We prefer to rely on the same, larger database to explore additional hypotheses, and continue to build our database to include additional metrics of health and sustainability.

Second, we agree with the importance of TL in assessing fisheries sustainability, although this relationship is more complex than previously thought. While it is commonly assumed that high TL species are more exploited than low TL species, recent data suggest that more populations of low TL species have collapsed as compared with those for large predators (Pinsky *et al.* 2011). That said, there is strong evidence that both body size and TL are positively related to Hg concentration (Cutshall *et al.* 1978; Cabana and Rasmussen 1994; Wiener and Spry 1996; Hammerschmidt and Fitzgerald 2006; Burger and Gochfeld 2011; Figure 1 in Tlusty's letter). In our paper, we discuss TL and body size as possible mechanisms linking Hg and sustainability, explaining that carnivorous species tend to eat higher on the food chain, live longer, and hence biomagnify and bioaccumulate methylmercury. Our study was not

designed to test for the effects of body size and TL, and we do not claim to have found direct evidence for their effects. Instead, we offer them as likely possibilities explaining the identified relationship between Hg and sustainability.

Although Tlusty's TL dataset was unavailable, the simple 1–4 scale listed on <http://ocean.nationalgeographic.com/ocean/take-action/impact-of-seafood/#/marine-food-chain/> appears oversimplified given the wide range of ecological niches filled by common seafood items found in TL2–TL4. It is also unclear why Tlusty excluded body size from his analysis, since this can be more strongly related to Hg concentration than TL (Burger and Gochfeld 2011). We recognize that including TL and body size in the analysis is inherently difficult because these factors are variable within a species, probably obtained from different datasets, and values are not based on the same stock.

One sector where Tlusty did not find evidence supporting a link between TL, Hg, and sustainability was with farmed fish. Indeed, there are numerous challenges in quantifying TL for farmed fish. Is TL based on their typical trophic role in the wild, even though they are eating from a very different, artificial food chain? Did Tlusty rely on one TL per species or different TLs for farmed versus wild fish? The former would clearly introduce additional bias into the results for farmed fish. Such questions regarding how to compare wild and farmed stocks are challenging but important, given that farmed fish comprise an increasingly larger share of the market.

We encourage continued discussion about effective ways for consumers to make informed decisions about their seafood consumption. While we appreciate Tlusty's suggestion that we focus on TL to simplify seafood awareness initiatives, we question the relevance of this concept to the average seafood consumer. As stated in

our paper, consumers interpret sustainability in many ways, so it is important to search for associations that span multiple interpretations. In our view, our paper's original message is unchanged – by choosing sustainable species, you are also likely choosing healthier options, which, incidentally are also generally low TL species. However, both sustainability and TL have complex meanings, and sustainability status for different seafood items may change over time. Additional analyses should examine the extent to which other, simpler variables – such as body size – are associated with both health and sustainability metrics.

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Burger J and Gochfeld M. 2011. Mercury and selenium levels in 19 species of saltwater fish from New Jersey as a function of species, size, and season. *Sci Total Environ* **409**: 1418–29.

Cabana G and Rasmussen JB. 1994. Modeling food-chain structure and contaminant bioaccumulation using stable nitrogen isotopes. *Nature* **372**: 255–57.

Cutshall NH, Naidu JR, and Pearcy WG. 1978. Mercury concentrations in Pacific hake, *Merluccius productus* (Ayres), as a function of length and latitude. *Science* **200**: 1489–91.

Hammerschmidt CR and Fitzgerald WF. 2006. Bioaccumulation and trophic transfer of methylmercury in Long Island Sound. *Arch Environ Con Tox* **51**: 416–24.

Pinsky ML, Jensen OP, Ricard D, and Palumbi SR. 2011. Unexpected patterns of fisheries collapse in the world's oceans. *P Natl Acad Sci USA* **108**: 8317–22.

Wiener JG and Spry DJ. 1996. Toxicological significance of mercury in freshwater fish. In: Beyer W, Heinz GH, and Redmon-Norwood AW (Eds). *Environmental contaminants in wildlife: interpreting tissue concentrations*. Boca Raton, FL: Lewis Publishers.

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